Local habitat and landscape influence predation of bird nests on afforested Mediterranean cropland

SÁNCHEZ-OLIVER, J.S. [a], REY BENAYAS, J.M.[a,*] and CARRASCAL, L.M.[b]

[a]: Dpto. de Ciencias de la Vida, Edificio de Ciencias, Universidad de Alcalá, 28871 Alcalá de Henares, Spain. Tel. +34 91 8856408, Fax +34 91 8854929 (e-mail: juansalvador.sanchez@uah.es; josem.rey@uah.es)

[b]: Dpto. Biogeografía y Cambio Global, Museo Nacional de Ciencias Naturales, CSIC, José Gutiérrez Abascal, 2, 28006 Madrid, Spain (e-mail: mcnc152@mncn.csic.es)

*: Corresponding author during the review process; e-mail: josem.rey@uah.es

ABSTRACT

Afforestation programs such as the one promoted by the EU Common Agrarian Policy have contributed to spread tree plantations on former cropland. Nevertheless these afforestations may cause severe damage to open habitat species, especially birds of high conservation value. We investigated predation of artificial bird nests at young tree plantations and at the open farmland habitat adjacent to the tree plantations in central Spain. Predation rates were very high at both tree plantations (95.6%) and open farmland habitat (94.2%) after two and three week exposure. Plantation edge/area ratio and development of the tree canopy decreased predation rates and plantation area and magpie (*Pica pica*) abundance increased predation rates within tree plantations, which were also affected by land use types around plantations. The area of nearby tree plantations (positive effect), distance to the tree plantation edge (negative effect), and habitat type (mainly attributable to the location of nests in vineyards) explained predation rates at open farmland habitat. We conclude that predation rates on artificial nests were particularly high and rapid at or nearby large plantations, with high numbers of magpies and low tree development, and located in homogenous landscapes dominated by herbaceous crops and pastures with no remnants of semi-natural woody vegetation. Landscape planning should not favour tree plantations as the ones studied here in Mediterranean agricultural areas that are highly valuable for ground-nesting bird species.

Keywords: Artificial nests, Farmland habitat, Land use types, Magpie abundance, Pine plantations

1. Introduction

A significant amount of abandoned cropland, low productive cropland and pastureland has been converted into tree plantations in the last few decades, and ca. 7% of the total forest land in the world are tree plantations at present (FAO, 2011). Different afforestation programs have contributed to the spread of such tree plantations at the regional level. Thus, the Common Agricultural Policy (CAP) has favoured the conversion of farmland into tree plantations in the European Union since 1992 by means of a scheme of aid for forestry measures in agriculture (EEC Council Regulation No. 2080/92), which has resulted in the afforestation of > 8 million ha to date (European Commission, 2013a, 2013b). Further, afforested cropland is expected to increase in the near future in countries such as Spain due to subsidies to afforestation of extirpated vineyards (Spanish Agrarian Guarantee Fund, 2012). This afforestation program pursues both societal and environmental benefits, including control of erosion, prevention of desertification, regulation of the water regime, and increasing the fixation rate of carbon dioxide. However, whereas tree plantations provide a number of benefits (Rey Benayas et al., 2007), they may have noticeable effects on biological communities, as it has been exemplarily shown with birds (Shochat et al., 2001; Santos et al., 2006; Bremer and Farley, 2010; Felton et al., 2010; Lindenmayer et al., 2010; Rey Benayas et al., 2010).

Agro-ecosystems are important for maintenance of bird diversity in Europe, especially for species of conservation concern (BirdLife International, 2004). The Directorate-General for Agriculture and Rural Development (2012), using *the common farmland bird index* as “a barometer of change for the biodiversity of agricultural land in Europe", shows a decline in these bird populations of ca. 20% between 1990 and 2008 (see also Gregory et al., 2005; Butler et al., 2010; Guerrero et al., 2012). Cropland afforestations in southern Europe are mostly based on coniferous species such as *Pinus halepensis* and *P. pinaster*, and are an example of novel and hybrid ecosystems sensu Hobbs et al. (2009). These plantations may cause damage to open habitat species, especially birds, by replacing high quality open farmland habitat and increasing risk of predation (Díaz et al., 1998; Cresswell, 2008; Reino et al., 2009). Predation has both direct and indirect effects on bird populations (Batáry and Báldi, 2004), the latter related to the avoidance of use of habitats that are perceived as risky (Murcia, 1995) or fecundity reduction (Bonnington et al., 2013). Besides hindering the persistence of established ground-nesting bird populations, predation may impede the colonization of the new afforested habitat by bird species (Murcia, 1995; Lindenmayer and Fischer, 2006).

Tree plantations act as sources of generalist predators of various types, including rodents, lagomorphs, feral cats, dogs, and corvids (Andren, 1992; Pita et al., 2009; Reino et al., 2010; Suvorov et al., 2012). These generalist predators usually have very low densities at treeless open habitats, but thrive in mosaic habitat landscapes where they exhibit an exploratory behaviour (Andren, 1992; Pita et al., 2009; Reino et al., 2010). Particularly, predation by corvids is enhanced in humanized landscapes where they attain high densities (Jokimaki et al., 2000; Newson et al., 2010), and Salek (2004) experimentally showed that the presence of magpie (*Pica pica*) nests increased predation rates on bird eggs. Accordingly, Castilla et al. (2007) attributed in part the relatively low predation on Red-legged Partridge (*Alectoris rufa*) eggs at Mediterranean fallow fields to the low presence of magpies due to their capture by humans. Magpies are strongly attracted by trees in deforested landscapes for nesting, and this phenomenon is highly noticeable at relatively small and isolated tree plantations in Mediterranean cropland afforestations.

This study aimed to investigate the predation of bird eggs set on artificial nests at young (< 20 yr) tree plantations established on former cropland and at the open habitat adjacent to such tree plantations in a farmland and woodland Mediterranean mosaic. We hypothesized that nest predation will be affected by both (1) the features of local breeding habitat and (2) the features of landscape –namely proportion of land use types- surrounding local habitat. At tree plantations, we predicted that (i) a reduced area and a high edge-area ratio will favour permeability to predators and hence increase nest predation rates and (ii) magpie abundance and predation rate will be positively correlated. At open farmland habitat adjacent to tree plantations, we predict that predation rates will be influenced by (i) plantation area (positive effect), (ii) distance from plantation (negative) and (iii) magpie abundance (positive).

Our experimental study sheds light on the risk of nest predation at Mediterranean landscapes that have been subjected to afforestation projects of former cropland, and provides insights for impact assessment and management of such projects at the local habitat and landscape scales.

2. Methods

2.1. Study area

Field work was carried out in afforested cropland and open farmland located in Campo de Montiel (La Mancha natural region, southern Spanish plateau, 38º41’53”N, 2º51'54"W, **Figure S1 in Supplemental Material**). The study area spreads on ca. 440 km^2 with altitude ranging between 690 and 793 m a.s.l. The climate is continental Mediterranean with dry and hot summers and cold winters. Mean annual temperature and total annual precipitation in the area during the last 30 years were 13.7 °C and 390 mm, respectively (Agencia Española de Meteorología, 2012). These figures were 16.6°C and 359.9 mm in 2011 and 15.8ºC and 362.9 mm in 2012, when our nest predation experiments took place (Junta de Castilla-La Mancha, 2013).

The area is a representative mosaic of different crops, pastures and semi-natural or planted woody vegetation that are characteristic of large areas in Mediterranean landscapes. Croplands were mostly occupied by herbaceous crops (wheat and barley) and permanent woody crops (olive groves and vineyards). Natural vegetation consisted of holm oak (*Quercus rotundifolia L.*) woodland and riparian forests that have been mostly extirpated from this region. Until 1992, woodland cover was restricted to open holm oak parklands, usually grazed by sheep and goats. Major land use changes in the last 20 years are the abandonment of herbaceous cropland and vineyard extirpation and their subsequent afforestation with the native Aleppo pine (*Pinus halepensis* Mill.) alone or mixed with holm oak and (*Retama sphaerocarpa* (L.) Boiss) (Rey Benayas et al., 2010). These tree plantations are noticeably dominated by pines as they establish better and grow faster than the other planted species.

2.2. Selection of tree plantations for predation experiments

The constraints associated with ~~to~~ each habitat type, namely tree plantations and open farmland adjacent to tree plantations, prevented homogeneous experimental designs and sampling methods, and consequently data from the different experiments were not directly analysed together (see below). Thus, we run two independent experiments of bird nest predation, (1) at tree plantations and (2) on open farmland. First, all tree plantations in the study area were located using both orto-photos (Geographic Information System of Farming Land, 2010; hereafter SigPac) and Google Earth®, and were later verified in the field. We found 99 tree plantations on former cropland that took place in 1992 or later. Only tree plantations > 0.78 ha were selected for the predation experiments to take advantage of bird survey plots of this size in the study area. In addition, a target tree plantation for the experiment on adjacent farmland had to be placed at least 2-km away from another plantation to avoid that experimental nests associated with a given tree plantation were affected by another tree plantation. Following these criteria, we finally selected 30 tree plantations for the experiment at tree plantations and 38 tree plantations for the experiment on farmland adjacent to the tree plantations, with 20 plantations that were used in both experiments (**Figure S1 in Supplemental Material**).

2.3. Survey of magpie abundance

We recorded the abundance of magpie as a potential nest predator in the studied tree plantations and open farmland habitat adjacent to such plantations. At every tree plantation, magpies were surveyed using point-count stations (Bibby et al., 2000) lasting 10 minutes in May 2011. The point-counts were located at the centre of each tree plantation. All auditory and visual contacts were recorded, but only those within a 50 m radius (0.78 ha; **Figure S2 in Supplemental Material**) were used in subsequent analyses, in order to increase the probability of detection. Every point-count station was surveyed by two censuses in different days, one within the first 4 h in the morning and another in the afternoon beginning 3 h before sunset. We used the average of the two counts as a measure of magpie abundance. The same trained person conducted all the censuses (JSS-O) on nearly windless (wind speed <3 m s^{-1}) and rainless days.

The open farmland habitat adjacent to 38 tree plantations was also surveyed for magpie abundance by means of one line transect of 400-m length and 200-m width in may 2012 (**Figure S2 in Supplemental Material**). Again, all censuses were conducted by the same well trained field ornithologist (JSS-O) on windless (wind speed < 3 m s-1) and rainless days. We employed two different census methods and years for sampling magpie relative abundance according to the limitations imposed by the size of pine plantations, where transects were not possible due to their small area. Nevertheless, this is not a concern in this study as the aim is not to compare magpie abundance inside *vs* outside plantations, but to relate the relative abundance of magpies to nest predation within plantations and outside plantations, separately.

Other corvid species were disregarded as key predators of artificial nests because they were very scarce in the study area (the Carrion Crow, *Corvus corone*, was detected at only one open farmland adjacent to tree plantations, and other species such as the Jay, *Garrulus glandarius*, or the Raven, *C. corax*, were not observed in the study area).

2.4. Nest predation experiments

The two nest predation experiments used quail (*Coturnix coturnix*) eggs that were layed on exposed artificial wicker nests (two eggs at each artificial nest; see below details on egg placement). All eggs had the same origin (i.e., supplier), were washed and then dried at air temperature before being used for the field experiments (Vander Haegen and DeGraaf, 1996; Conner and Perkins, 2003; Piper and Catterall, 2004), and were handled with gloved hands to minimize human scent (Whelan et al., 1994).

The artificial nests at tree plantations and on open fields near plantations were not placed on the same date due to limitations inherent to the organization of the field work, which included a number of tasks, and considering the timing of agricultural activities in the study area (e.g., ploughing). Nevertheless, the data for the two experiments were analysed separately and were never directly compared.

We considered an artificial nest as predated when the eggs were either absent or damaged, excluding from analyses those artificial nests that were ploughed or trampled (42 and 7, respectively, on open farmland and neither at tree plantations). Types of predators could not be distinguished for the eggs that were removed from the artificial nests which, in turn, were most of them (see Results). Nevertheless, unidentified predation events were probably attributable to small corvids (Schaefer, 2004) such as magpies considering their ability to store large items of food and to steal and remove eggs from nests (Henty, 1975; Groom, 1993; Perrins, 1998). We were able to distinguish predation by rodents (by their characteristic bites and, sometimes, faeces) and by corvids (by their characteristic pecks) from some fresh egg remains, whereas for the largest part of predated eggs with fresh remains we could not distinguish the source of predation. However, this issue is not a problem for the aims of this study since we were interested in the effects of tree plantations on overall predation risk rather than in the identification of predator assemblages.

Experiment 1.- Predation at the tree plantations. This experiment was run at 30 plantations in the spring of 2011, which averaged 5.6 ± 7.2 ha and ranged between 1.5

and 36.5 ha. The artificial nests with two quail eggs each were placed at two different positions (i.e. one nest on the ground and another nest on pine branches) at 25-m intervals along an *a priori* line spanning from the edge (0 m) to the centre of the plantation (**Figure S2 in Supplemental Material**), in May 22-25. The height above the ground for those nests located on branches was estimated using a measuring tape. The line where both on-ground and on-branches artificial nests were placed covered at least 50 m (i.e. three nest locations at 0, 25, and 50 m from the plantation edge), whereas the maximum length of that line from the plantation edge was 225 m that included ten nest locations (average was 70.8 m and sd = 38.9). Total sample size was 230 nests, 115 located on the ground and 115 located on branches. We visited the nests in two occasions, 7-9 days (May 31 and June 1) and 15-18 days after they were placed (June 9-11), counting the number of eggs that had been removed. Artificial nests were not checked more often in order to reduce the effect of the observer on predation and to preserve nest concealment (e.g., Major and Kendal, 1996).

Experiment 2.- Predation on open farmland adjacent to tree plantations. This experiment was run at 38 plantations in the spring of 2012. Each artificial wicker nest was baited with two treated quail eggs (see above) and was placed on the ground along an *a priori* 300-m line; this line spanned at 25-m intervals from the plantation edge (0 m) until 150 m away from such edge, and then at 50-m intervals until 300 m (i.e., nine nests at 0, 25, 50, 75, 100, 150, 200, 250, and 300 m; **Figure S2 in Supplemental Material**). The artificial nests were placed on May 4-9. Total sample size was 342 nests. We took note of the habitat type where each nest was situated, considering five habitat categories (olive groves, vineyard, abandoned cropland and pastures, semi-natural woody vegetation, and dry herbaceous cropland). We checked the nests for egg predation in two occasions (in May 15-22 and in May 27-June 1, 11-14 days and 21-23 days after the nests were placed), following the same protocol presented in Experiment 1.

2.5. Local habitat and landscape features

In each of the 46 tree plantations where experiments 1 and 2 took place, we characterized variables related to vegetation structure, area, edge/area ratio, and landscape surrounding the tree plantation (**Table S1 in Supplemental Material)**. Vegetation structure at each surveyed plantation was characterized in one 25-m radius plot (**Figure S2 in Supplemental Material**). We directly measured or estimated by eye, after previous training, the following structural features of the vegetation: percentage cover of chamaephytes, shrubs and trees, average height of chamaephytes, shrubs and trees, and number of trunks <5, 5-10, 10–20, 20–40 and >40cm in diameter at breast height (dbh). Additionally, we estimated percentage cover of herbs and bare soil and measured the average height of the herb layer in one concentric 10-m radius plots within the 25-m radius plot (**Figure S2 in Supplemental Material**) due to perceptual limitations when carrying out visual estimations. Vegetation structure was sampled by the same observer (JSS-O) to avoid inter-personal bias in vegetation measurements. We also measured area and edge/area ratio using ArcGIS 10.0 (ESRI Inc.). Edge/area ratio was calculated as the quotient between the length of the edge (in m) and the square root of the plantation area (in m^2).

Land use types were identified by means of land use layers taken from SigPac (see source above) and were analysed with ArcGIS 10.0. We distinguished 14 land use types: streams, rivers and lagoons, roads and rural tracks, urban areas and scattered buildings, semi-natural woodland, dried-fruit orchards, orchards, waste lands, olive grove, pastures with scattered trees, scrubland, pasture land, dry herbaceous cropland, vineyard, and vineyard with olive trees. To characterize landscape surrounding the tree plantations for Experiment 1, the percentage of area of each land use types was obtained in 1-km buffer-rings centred at each forest plantation (**Figure S2 in Supplemental Material**). To characterize landscape for Experiment 2, the proportion of land use types was measured as above at 600 m x 600 m squares that included the 300-m transects in farmland habitat were the artificial nests were set (**Figure S2 in Supplemental Material**).

2.6. Statistical analysis

The two experiments of nest predation were analysed independently. We used predation incidence obtained from the first checking date as most artificial nests were predated within the first 7-14 days after they were placed on the field (see Results).

We looked at the correlation among the independent variables of our models (see below). Most correlations were not significant. Moreover, the VIF figures (variance inflation factor) for predictors in the analyses were very low (<1.75). Particularly, the shared variance between magpie abundance and other explanatory variables was usually very low (as measured by the coefficient of determination R^2): (a) Within tree plantations: log area 0.03; edge/area ratio <0.001; PC1 vegetation 0.14; PC2 vegetation 0.04; and PC1 land use 0.08; (b) On open farmland habitat: log area 0.09; pine height <0.001; and PC1 land use 0.03.

Two statistical approaches were carried out for each experiment. First, we analysed the predation of each individual nest using a binomial response variable (predated-1, non-predated-0) by means of a Generalized Linear Mixed Effects Model, with the study location (the plantation or the farmed fields outside the plantation) as a random factor and the position of the nests as fixed effects. Additionally, we analysed global predation rates at the tree plantations and on open farmland by means of a generalized Poisson regression model. This model considered the whole sample of artificial nests at each location. Predictor variables described the characteristics of the plantations, the density of the magpie and the landscape structure around each study location.

Experiment 1.- A Generalized Linear Mixed Model was applied to the predation of each individual nest at tree plantations using a binomial response variable (predated-1, non-predated-0; logit link function). The plantation was included as a random factor and the position of the nests were the fixed effects (distance of each artificial nest to the plantation edge and height of artificial nests above the ground). The continuous predictor variables were standardized to mean = 0 and sd = 1 in order to obtain standardized regression coefficients. Statistical significance was estimated using a robust approach with quasi-ML standard errors (Lindsey, 2004) after correcting for overdispersion ($\phi = 0.82$).

We also used a Generalized Linear Model based on a Poisson distribution (with the log-link function) for analysing the number of predated nests, with the total number of artificial nests placed at each plantation as an offset. This model was applied to analyse the effects of six predictor variables, namely tree plantation area (log-transformed), plantation edge/area ratio, magpie abundance, two components related to vegetation structure, and a principal component related to landscape features (see below). Statistical significance of the standardized regression coefficients of the predictor variables was estimated using a robust approach with quasi-ML standard errors after correcting for overdispersion ($\phi = 0.72$).

We performed two different principal components analyses (PCA), one on vegetation structure variables within tree plantations and another on land use types surrounding the plantations, to obtain synthetic and independent environmental gradients that may affect nest predation.

Experiment 2.- A Generalized Linear Mixed Model was applied to the predation of each individual nest on farmland habitat using a binomial response variable (predated-1, non-predated-0; logit link function). The plantation was included as a random factor and the position of the nests were the fixed effects (distance of each artificial nest to the nearest plantation edge and a factor describing the location in six different habitat categories). The continuous predictor variables were standardized to obtain standardized regression coefficients. Statistical significance was estimated using a robust approach with quasi-ML standard errors after correcting for overdispersion ($\phi = 0.36$).

Additionally, we used another Generalized Linear Model based on a Poisson distribution (with the log-link function). The response variable was the number of predated nests placed at each transect, with the total number of artificial nests as an offset. Predictor variables were: area of the nearby tree plantation (log-transformed), average tree height of the nearest plantation, magpie abundance on the farmed field transect, and the principal component related to landscape features (see below). The continuous predictor variables were standardized in order to obtain standardized regression coefficients. Statistical significance of the standardized regression coefficients of the predictor variables was estimated using a robust approach with quasi-ML standard errors after correcting for overdispersion (ϕ = 0.43).

For this experiment we carried out another PCA on land use type categories measured at 600 m x 600 m squares.

Out of the 342 nests placed in total for the experiment, 40 nests were not found, seven were trampled, and 42 were located on cropland fields that were ploughed. All artificial nests at five out of the 38 tree plantations that were initially selected for Experiment 2 were lost due to ploughing or trampling.

Statistical analyses were carried out using GRETL 1.9.14 (Cottrell and Lucchetti, 2007) and STATISTICA 10 (StatSoft, 2011).

3. Results

3.1. Dominant environmental gradients

The first component on vegetation structure variables within tree plantations (51.2% of total variance) defined a gradient of increasing development of the tree canopy, as it opposed tree cover (factor loading = 0.897), tree height (0.816) and number of trunks >5 cm in dbh (0.852) to shrub height (-0.724) and cover (-0.523) and herb height (-0.656). The second component on vegetation structure variables (20.1% of the total variance) was associated with the development of the shrub layer; it opposed shrub cover (0.727) and height (0.602) to herb cover (-0.611). The first component on land use around tree plantations (36.3% of the total variance) opposed olive groves (0.964) and semi-natural woodland (0.718) to roads and rural tracks (-0.842), vineyard (-0.766) and dry herbaceous cropland (-0.637).

For land use type categories measured at 600 m x 600 m squares on open farmland habitat, the first component (14.8% of the total variance) opposed semi-natural woodland (0.644) and pastures with scattered trees (0.626) to dry herbaceous cropland (-0.715).

3.2. Predation rates and magpie abundance

Overall predation rates were very high at both the tree plantations and adjacent open farmland (**Figure 1**). 81.2% and 88.4% of the predated artificial nests were observed at tree plantations and on open farmland habitat, respectively, by the first counting, one to two weeks after being set. Only 4.4% and 5.8% of artificial nests at tree plantations and on open farmland habitat, respectively, were left un-predated two to three weeks after the start of the experiment (**Figure 1**).

All artificial nests at 12 (40%) tree plantations were predated by the first counting, and all artificial nests were left un-predated at only one tree plantation. On open farmland habitat, all artificial nests were predated in 21 (58.3%) transects by the first counting. The maximum number of artificial nests left un-predated in a transect by the first counting was 85.71%.

Of the total nests, 74.2% at tree plantations and 79.2% on open farmland were removed and, consequently, their source of predation is unknown. Predation by rodents at tree plantations and on open farmland were, respectively, 1.7% and 2.3%, whereas

5.2% and 6.9%, respectively, showed evidence of predation by corvids, namely magpie as the nearly exclusive corvid species present around and in plantations.

Mean magpie abundance at the 30 tree plantations was 1.37 birds per ha (sd = 1.87, range = 0-6.41), whereas it averaged 0.11 birds ha^{-1} (sd = 0.18, range =0-0.63, n=38) at open farmland near tree plantations.

3.3. Nest predation at tree plantations

The Generalized Linear Mixed Model analysing the predation probability of artificial nests at tree plantations showed substantial differences among plantations, but distance of artificial nests to the plantation edge and height of nests above the ground did not show any significant effect on nest predation (**Table 1**).

The Generalized Linear Model (Poisson distribution) of the number of predated nests at tree plantations, using the total number of artificial nests placed at each plantation as an offset **(Table 2)**, revealed positive effects of tree plantation area and magpie abundance (**Figure 2**), and negative effects of edge/area ratio, development of the tree canopy (first PC of vegetation structure variables), and relative amount of tree crops and woodland in the landscape (first PC of land use type variables).

3.4. Nest predation on open farmland adjacent to tree plantations

The Generalized Linear Mixed Model analysing nest predation of individual nests on open farmland adjacent to tree plantations resulted in significant effects of the three predictors (**Table 3**). There were important differences among the 33 open farmland sites adjacent to tree plantations (random factor). Distance of nests to the nearest edge of tree plantations had a negative effect on predation risk (i.e., lower predation risk at longer distances from plantations). The habitat type where artificial nests were placed had also a significant effect, mainly attributable to the location of nests in vineyards that increased the probability of predation.

The Generalized Linear Model (Poisson distribution) of the number of predated nests on open farmland adjacent to tree plantations, using the total number of artificial nests placed outside plantations as an offset, showed only a significant effect of the nearby plantation area, global predation risk on open farmland being higher around larger tree plantations. Mean height of nearby tree plantations, magpie abundance and the relative amount of tree crops and woodland in the landscape (first PC of land use variables) did not show any significant effect on nest predation **(Table 4)**.

The area of the tree plantations for predated (n = 224) and non-predated (n = 29) artificial nests were (mean $\pm$ se) 6.4 ± 0.48 ha and 3.1 ± 0.24 ha, respectively. Predated and non-predated artificial nests were on average at a distance of 121.0 ± 6.38 m and 144.0 ± 20.75 from the tree plantations, respectively, and the modal values corresponded to a distance of 50 m for predated nests and 300 m for non-predated nests.

4. Discussion

Overall, we found that predation of artificial bird nests at young tree plantations established on former cropland and at adjacent open farmland habitat in a Mediterranean mosaic located in central Spain was (1) very high at both habitats, (2) influenced by local habitat features, and (3) influenced by landscape context. However, we obtained different results for specific variables that were hypothesized to affect predation rates inside and outside the investigated tree plantations (i.e. area, edge/area ratio, distance to edge, and magpie abundance).

The use of artificial nests to test predation rates is controversial due to factors that are not controlled with respect to real nests, and several studies have demonstrated that artificial nests do not estimate nest predation rates on natural nests precisely (Burke et al., 2004; Faaborg, 2004; Thompson and Burhans, 2004; Villard and Part, 2004). Thus, artificial nests can reveal where birds would never choose to nest as opposed to where their nests would suffer relatively high predation rates. Also, nest predation is only one of demographic parameters and thus this study provides only a partial view of the ecological relationships in the studied landscape.

4.1. High predation rates on artificial nests

Nest predation was very high and quick at both the tree plantations and adjacent open farmland habitat (>80% in less than two weeks after the start of our experiments). These rates are among the highest reported in the scientific literature (data and references in **Table 5**). Previous published figures of nest predation rates at tree plantations average 59.5% with a range of 23-94% (**Table 5a**). Similarly, nest predation rates for natural forest fragments are usually high (mean = 66.4%, range = 38.9-88.0%; **Table 5b**) but lower than in our tree plantations (95.6%). Other studies that have assessed nest predation rates at open habitat adjacent to tree plantations or natural forest fragments reported figures that average 60.0% (range = 13.7-100%; **Table 3c**), which are substantially lower than our 94.2% predation rate. However, comparisons of these figures with the figures obtained in our study should be cautious due to the different experimental designs across studies. In an experiment that used eggs of red-legged partridge located at Holm oak woodland patches in Central Spain, Castilla et al. (2007) reported a predation rate of 38.9% after a 2-week exposure.

We attribute the high predation rates in our study to the following three phenomena. First, our tree plantations were overall very small (mean size of 5.7±6.7 ha), which make nests easily accessible to predators in general even at the largest plantations (Ford et al., 2001; Chalfoun et al., 2002). Second, they were located in an agricultural and highly humanized landscape, which may favor predation by a number of animals such as rodents, hares, feral cats and dogs (Danielson et al., 1997; Jokimäki et al., 2005; Pangau-Adam et al., 2006). Also, short vegetation −such as that in the fields surrounding the studied plantations- is usually associated with very high predation rates (Beja et al., 2014). And third, they were a very attractive habitat for magpies, a documented powerful nest predator (Andren, 1992; Roos, 2004; Suvorov et al., 2012) that was particularly abundant in our study area (Sánchez-Oliver et al., 2013).

4.2. Factors affecting predation rates

Nest predation at tree plantations increased with larger plantation area and abundance of magpies and with a lower edge-area ratio and development of the tree layer, whereas nest predation on open farmland habitat was higher if nearby tree plantations were of large area and nests were located at closer distances from the plantations.

The small size and homogeneity of the studied tree plantations and the high predation rates explain why distance to edge and average height above the ground of individual nests did not have any effect on predation rates at tree plantations. However, a shorter distance to edge of the tree plantation may enhance predation on the open farmland habitat because nests are closer to the source of predators (Batáry and Báldi, 2004; Reino et al., 2010) such as magpies. Lack of association between magpie abundance and nest predation on open farmland makes unclear if magpies are or not a major predator in open habitats, an issue that should be tested by using cameras to identify the actual predators.

Predator identification in our experiments was relatively unsuccessful as the proportion of eggs that disappeared was high (>74%) and, unfortunately, egg shell observation did not provide enough information to determine the main sources of predation. However, we detected a positive relationship between nest predation rates and magpie abundance inside tree plantations, which points to relevance of nest predation caused by magpies. Magpies have a high capability of exploring relatively new habitats and are prone to nesting in the most developed plantations (> 3 m in height) that we surveyed in our mostly deforested study area (Carrascal et al., 2014). Andren (1992) found predation rates of bird nests in forest fragments by this corvid that ranged between 7.2% and 35.7%. As most of the studied tree plantations are of a rectangular shape, low edge-area ratios mean larger plantations, which may function as a refuge habitat and harbour a higher abundance of magpies and other generalist predators of bird nests such as domestic carnivores (Virgós et al., 2002; Barea-Azcón et al., 2006; Pita et al., 2009; Fandos et al., 2012).

The higher predation rates at tree plantations with lower tree development may be explained by the facts that these plantations are newer habitats that call more the attention of exploring predators (Virgós et al., 2002) than older plantations and, additionally, they are more open and thus artificial nests are more visible (Suvorov et al., 2012). The same influence of low vegetation cover can be related to the higher predation risk suffered by artificial nests located at vineyards outside tree plantations, a heavily anthropized habitat with no vegetation at ground level due to agricultural practices that eliminate the natural herbaceous layer that may compete for water and nutrients with grapevines.

Finally, we found significant landscape effects on nest predation at both the tree plantations and the surrounding open farmland habitat. Other studies have found relationships between landscape context and nest predation rates (Huhta et al., 1996; Bayne and Hobson, 1997). In our study, nest predation in tree plantations was higher in landscapes with higher proportion of herbaceous crops and pastures and lower proportion of woody crops and semi-natural woodlands. This finding supported by correlational evidence hints at the importance of semi-natural vegetation for conservation of ground-nesting birds in vast open farmed fields (Santos et al., 2006; Zuria et al., 2007; Ludwig et al., 2012). However, some open farmland birds may have strong negative reactions to woody habitats, either natural or planted, and they may also increase the abundance of generalist predators (Pita et al., 2009; Reino et al., 2009 and 2010). Deforested landscapes with a high proportion of herbaceous crops favour also the abundance of lagomorphs, which can predate on eggs (Reino et al., 2010). In general, tree plantations in open, deforested, and homogenous landscapes are better attractors and refuges of predators than tree plantations in more heterogeneous landscapes where there is more availability of habitat with trees (e.g. Andren, 1992).

5. Conclusion

Our experiments on predation rates at young afforestations of Mediterranean cropland and adjacent open farmland hint local habitat and landscape features that are indicators of predation risk for bird nests. We conclude that predation rates on artificial nests were particularly high and rapid at or nearby large plantations, with high numbers of magpies and low tree development, and located in homogenous landscapes dominated by herbaceous crops and pastures with no remnants of semi-natural woody vegetation. Thus, the studied tree plantations should not be favoured, and even be extirpated, in agricultural landscapes that are highly valuable for ground-nesting bird species and open farmland bird communities (Traba et al., 2006; Sánchez-Oliver et al., 2013). We recommend assessments of real nest predation risk following afforestation in agricultural landscapes to fully understand and, consequently, reduce its impacts on biodiversity, particularly on ground-nesting birds.

Acknowledgments

Projects from the Spanish Ministry of Science and Education (CGL2010-18312) and the Government of Madrid (S2009AMB-1783, REMEDINAL-2) are currently providing financial support for this body of research. JSS-O was supported by a contract from Fundación Internacional para la Restauración de Ecosistemas (www.fundacionfire.org). We are indebeted to two anonymous reviewers who improved the contents and presentation of a previous version of this manuscript. We also thank Claire Jasinski for improving the English of the first draft of the manuscript. All authors of this manuscript do not have any conflict of interest to declare.

References

Agencia Española de Meteorología, 2012. Agencia Española de Meteorología [WWW Document]. URL http://www.aemet.es/

Andren, H., 1992. Corvid density and nest predation in relation to forest fragmentation - A lansdcape perspective. Ecology 73, 794–804.

Arango-Vélez, N., Kattan, G., 1997. Effects of forest fragmentation on experimental nest predation in Andean cloud forest. Biol. Conserv. 3207, 137–143.

Barea-Azcón, J.M., Virgós, E., Ballesteros-Duperón, E., Moleón, M., Chirosa, M., 2006. Surveying carnivores at large spatial scales: a comparison of four broad-applied methods. Biodivers. Conserv. 16, 1213–1230.

Batáry, P., Báldi, A., 2004. Evidence of an Edge Effect on Avian Nest Success. Conserv. Biol. 18, 389–400.

Bayne, E.M., Hobson, K.A., 1997. Comparing the Effects of Landscape Fragmentation by Forestry and Agriculture on Predation of Artificial Nests. Conserv. Biol. 11, 1418–1429.

Beja, P., Schindler, S., Santana, J., Porto, M., Morgado, R., Moreira, F., Pita, R., Mira, A., Reino, L., 2014, Predators and livestock reduce bird nest survival in intensive Mediterranean farmland. Eur. J. Wildlife Res. 10.1007/s10344-013-0773-0.

Bibby, C., Burgess, N.D., Hill, D.A., Mustoe, S.H., 2000. Bird Census Techniques, 2nd ed. Academic Press, London.

BirdLife International, 2004. Birds in Europe: population estimates, trends and conservation status. BirdLife International, Cambridge, U.K.

Bonnington, C., Gaston, K.J., Evans, K.L., 2013. Fearing the feline: domestic cats reduce avian fecundity through trait-mediated indirect effects that increase nest predation by other species. J. Appl. Ecol. 50, 15–24.

Bremer, L.L., Farley, K.A., 2010. Does plantation forestry restore biodiversity or create green deserts? A synthesis of the effects of land-use transitions on plant species richness. Biodivers. Conserv. 19, 3893–3915.

Burke, D.M., Elliott, K., Moore, L., Dunford, W., Nol, E., Phillips, J., Holmes, S., Freemark, K., 2004. Patterns of Nest Predation on Artificial and Natural Nests in Forests. Conserv. Biol. 18, 381–388.

Butler, S.J., Boccaccio, L., Gregory, R.D., Vorisek, P., Norris, K., 2010. Quantifying the impact of land-use change to European farmland bird populations. Agric. Ecosyst. Environ. 137, 348–357.

Carignan, V., Villard, M.-A., 2002. Effects of variations in micro-mammal abundance on artificial nest predation in conifer plantations and adjoining deciduous forests. For. Ecol. Manage. 157, 255–265.

Carrascal, L.M., Galván, I., Sánchez-Oliver, J.S., Rey-Benayas, J.M., 2014. Regional distribution predicts bird occurrence in Mediterranean cropland afforestations. Ecol. Res. 29, 203–211.

Castilla, A.M., Dhondt, A.A., Díaz-Uriarte, R., Westmoreland, D., 2007. Predation in Ground-Nesting Birds: an Experimental Study Using Natural Egg-Color. Avian Conserv. Ecol. 2, 2.

Chalfoun, A.D., Thompson III, F.R., Ratnaswamy, M.J., 2002. Nest predators and fragmentation: a review and meta-analysis. Conserv. Biol. 16, 306–318.

Conner, L.M., Perkins, M.W., 2003. Nest predator use of food plots within a forest matrix: an experiment using artificial nests. For. Ecol. Manage. 179, 223–229.

Cottrell, A., Lucchetti, R., 2007. Gretl User's Guide.

Cresswell, W., 2008. Non□lethal effects of predation in birds. Ibis (Lond. 1859). 150, 3–17.

Danielson, W.R., DeGraaf, R.M., Fuller, T.K., 1997. Rural and suburban forest edges: effect on egg predators and nest predation rates. Landsc. Urban Plan. 38, 25–36.

Díaz, M., Carbonell, R., Santos, T., Tellería, J.L., 1998. Breeding bird communities in pine plantations of the Spanish plateaux: biogeography, landscape and vegetation effects. J. Appl. Ecol. 35, 562–574.

Directorate-General for Agriculture and Rural Development, 2012. Rural Development in the European Union. Statistical and Economic Information. Report 2012.

Estrada, A., Rivera, A., Coates-Estrada, R., 2002. Predation of artificial nests in a fragmented landscape in the tropical region of Los Tuxtlas, Mexico. Biol. Conserv. 106, 199–209.

European Commission, 2013a. Forestry measures under the common agricultural policy [WWW Document]. URL http://ec.europa.eu/agriculture/envir/report/en/forest_en/report.htm

European Commission, 2013b. EU agriculture - Statistical and economic information [WWW Document]. URL http://ec.europa.eu/agriculture/statistics/agricultural/index_en.htm

Faaborg, J., 2004. Truly Artificial Nest Studies. Conserv. Biol. 18, 369–370.

Fandos, G., Fernández-López, J., Tellería, J.L., 2012. Incursion of domestic carnivores around urban areas: a test in central Spain. Mammalia 76.

FAO, 2011. State of the World's Forests 2011. FAO, Rome.

Felton, A., Knight, E., Wood, J., Zammit, C., Lindenmayer, D.B., 2010. A meta-analysis of fauna and flora species richness and abundance in plantations and pasture lands. Biol. Conserv. 143, 545–554.

Ford, H., Barrett, G., Saunders, D., Recher, H., 2001. Why have birds in the woodlands of Southern Australia declined? Biol. Conserv. 97.

Geographic Information System of Farming Land, 2010. Geographic Information System of Farming Land [WWW Document]. URL http://pagina.jccm.es/agricul/sigpac.htm

Gregory, R.D., van Strien, A., Vorisek, P., Gmelig Meyling, A.W., Noble, D.G., Foppen, R.P.B., Gibbons, D.W., 2005. Developing indicators for European birds. Philos. Trans. R. Soc. Lond. B. Biol. Sci. 360, 269–88.

Groom, D.W., 1993. Magpie Pica pica predation on Blackbird Turdus merula nests in urban areas. Bird Study 40, 55–62.

Guerrero, I., Morales, M.B., Oñate, J.J., Geiger, F., Berendse, F., Snoo, G. De, Eggers, S., Pärt, T., Bengtsson, J., Clement, L.W., Weisser, W.W., Olszewski, A., Ceryngier, P., Hawro, V., Liira, J., Aavik, T., Fischer, C., Flohre, A., Thies, C., Tscharntke, T., 2012. Response of ground-nesting farmland birds to agricultural intensification across Europe: Landscape and field level management factors. Biol. Conserv. 152, 74–80.

Vander Haegen, W.M., DeGraaf, R.M., 1996. Predation rates on artificial nests in an industrial forest landscape. For. Ecol. Manage. 86, 171–179.

Henty, C.J., 1975. Feeding and food-hiding responses of jackdaws and magpies. Br. Birds 68, 463–466.

Hobbs, R.J., Higgs, E., Harris, J.A., 2009. Novel ecosystems: implications for conservation and restoration. Trends Ecol. Evol. 24, 599–605.

Huhta, E., Mappes, T., Jokimaki, J., 1996. Predation on artificial ground nests in relation to forest fragmentation, agricultural land and habitat structure. Ecography (Cop.). 19, 85–91.

Jokimaki, J., Huhta, E., Jokimäki, J., 2000. Artificial nest predation and abundance of birds along an urban gradient. Condor 102, 838.

Jokimäki, J., Kaisanlahti-Jokimäki, M.-L., Sorace, A., Fernández-Juricic, E., Rodriguez-Prieto, I., Jimenez, M.D., 2005. Evaluation of the "safe nesting zone" hypothesis across an urban gradient: a multi-scale study. Ecography (Cop.). 28, 59–70.

Junta de Castilla-La Mancha, 2013. Datos meteorológicos de Red de la Calidad del Aire de Castilla-La Mancha [WWW Document]. URL http://pagina.jccm.es/medioambiente/rvca/meteo.htm

Legendre, P., 1993. Spatial autocorrelation: trouble or new paradigm? Ecology 74, 1659–1673.

Lindenmayer, D.B., Fischer, J., 2006. Habitat fragmentation and landscape change: an ecological and conservation synthesis. CSIRO Publishing, Melbourne.

Lindenmayer, D.B., Knight, E.J., Crane, M.J., Montague-Drake, R., Michael, D.R., MacGregor, C.I., 2010. What makes an effective restoration planting for woodland birds? Biol. Conserv. 143, 289–301.

Lindsey, J.K., 2004. Introduction to Applied Statistics. A modelling Approach. Oxford University Press, Oxford.

Ludwig, M., Schlinkert, H., Holzschuh, A., Fischer, C., Scherber, C., Trnka, A., Tscharntke, T., Batáry, P., 2012. Landscape-moderated bird nest predation in hedges and forest edges. Acta Oecologica 45, 50–56.

Major, R.E., Kendal, C.E., 1996. The contribution of artificial nest experiments to understanding avian reproductive success: a review of methods and conclusions. Ibis (Lond. 1859). 138, 298–307.

Murcia, C., 1995. Edge effects in fragmented forests: implications for conservation. Trends Ecol. Evol. 10, 58–62.

Newson, S.E., Rexstad, E.A., Baillie, S.R., Buckland, S.T., Aebischer, N.J., 2010. Population change of avian predators and grey squirrels in England: is there evidence for an impact on avian prey populations? J. Appl. Ecol. 47, 244–252.

Pangau-Adam, M.Z., Waltert, M., Mühlenberg, M., 2006. Nest Predation Risk on Ground and Shrub Nests in Forest Margin Areas of Sulawesi, Indonesia. Biodivers. Conserv. 15, 4143–4158.

Pedersen, Å.Ø., Yoccoz, N.G., Ims, R. a., 2009. Spatial and temporal patterns of artificial nest predation in mountain birch forests fragmented by spruce plantations. Eur. J. Wildl. Res. 55, 371–384.

Perrins, C.M., 1998. The complete birds of the western Palearctic on CD-ROM, version 1.0. Oxford University Press, Oxford.

Piper, S.D., Catterall, C.P., 2004. Effects of edge type and nest height on predation of artificial nests within subtropical Australian eucalypt forests. For. Ecol. Manage. 203, 361–372.

Pita, R., Mira, A., Moreira, F., Morgado, R., Beja, P., 2009. Influence of landscape characteristics on carnivore diversity and abundance in Mediterranean farmland. Agric. Ecosyst. Environ. 132, 57–65.

Purger, J.J., Meszaros, L., Purger, D., 2004. Ground nesting in recultivated forest habitats a study with artificial nests. Acta Ornithol. 39, 141–145.

Reino, L., Beja, P., Osborne, P.E., Morgado, R., Fabião, A., Rotenberry, J.T., 2009. Distance to edges, edge contrast and landscape fragmentation: Interactions affecting farmland birds around forest plantations. Biol. Conserv. 142, 824–838.

Reino, L., Porto, M., Morgado, R., Carvalho, F., Mira, A., Beja, P., 2010. Does afforestation increase bird nest predation risk in surrounding farmland? For. Ecol. Manage. 260, 1359–1366.

Rey Benayas, J.M., Bullock, J.M., 2012. Restoration of Biodiversity and Ecosystem Services on Agricultural Land. Ecosystems 15, 883–899.

Rey Benayas, J.M., de la Montana, E., Pérez-Camacho, L., De la Cruz, M., Moreno, D., Parejo, J.L., Suárez-Seoane, S., Galván, I. 2010. Short-term dynamics and spatial pattern of a nocturnal bird assemblage inhabiting a Mediterranean agricultural mosaic. Ardeola 57: 303-320.

Rey Benayas, J.M., Galván, I., Carrascal, L.M., 2010. Differential effects of vegetation restoration in Mediterranean abandoned cropland by secondary succession and pine plantations on bird assemblages. For. Ecol. Manage. 260, 87–95.

Rey Benayas, J.M., Martins, A., Nicolau, J.M., Schulz, J.J., 2007. Abandonment of agricultural land: an overview of drivers and consequences. CAB Rev. Perspect. Agric. Vet. Sci. Nutr. Nat. Resour. 2, 1–14.

Roos, S., 2004. Nest Predation Processes and Farmland Birds: Habitat Selection and Population Dynamics of Predators and Prey.

Salek, M., 2004. The spatial pattern of the Black-billed Magpie, Pica pica to predation risk on dummy nests. Folia Zool. 53, 57-64.

Sánchez-Oliver, J.S., Rey Benayas, J.M., Carrascal, L.M., 2013. Differential effects of local habitat and landscape characteristics on bird communities in Mediterranean afforestations motivated by the EU Common Agrarian Policy. Eur. J. Wildl. Res. In press.

Santos, T., Tellería, J.L., 1992. Edge effects on nest predation in mediterranean fragmented forests. Biol. Conserv. 60, 1–5.

Santos, T., Tellería, J.L., Díaz, M., Carbonell, R., 2006. Evaluating the benefits of CAP reforms: Can afforestations restore bird diversity in Mediterranean Spain? Basic Appl. Ecol. 7, 483–495.

Schaefer, T., 2004. Video monitoring of shrub-nests reveals nest predators: Capsule Jays Garrulus glandarius are the most common predators, but carnivorous mammals and some other species also predate nests. Bird Study 51, 170–177.

Shochat, E., Abramsky, Z., Pinshow, B., 2001. Breeding bird species diversity in the Negev: Effects of scrub fragmentation by planted forests. J. Appl. Ecol. 38, 1135–1147.

Spanish Agrarian Guarantee Fund, 2012. Spanish Agrarian Guarantee Fund [WWW Document]. URL www.fega.es

StatSoft, 2011. Statistica 10 (data analysis software system).

Suvorov, P., Svobodová, J., Koubová, M., Dohnalová, L., 2012. Ground Nest Depredation by European Black-Billed Magpies Pica pica : An Experimental Study with Artificial Nests. Acta Ornithol. 47, 55–61.

Thompson, F.R., Burhans, D.E., 2004. Differences in Predators of Artificial and Real Songbird Nests: Evidence of Bias in Artificial Nest Studies. Conserv. Biol. 18, 373–380.

Traba, J., García de la Morena, E.L., Morales, M.B., Suárez, F., 2006. Determining high value areas for steppe birds in Spain: hot spots, complementarity and the efficiency of protected areas. Biodivers. Conserv. 16, 3255–3275.

Villard, M.-A., Part, T., 2004. Don’t Put All Your Eggs in Real Nests: a Sequel to Faaborg. Conserv. Biol. 18, 371–372.

Virgós, E., Tellería, J.L., Santos, T., 2002. A comparison on the response to forest fragmentation by medium-sized Iberian carnivores in central Spain. Biodivers. Conserv. 11, 1063–1079–1079.

Whelan, C., Dilger, M., Robson, D., Hallyn, N., Dilger, S., 1994. Effects of olfactory cues on artificial-nest experiments. Auk 111, 945–952.

Zuria, I., Gates, J.E., Castellanos, I., 2007. Artificial nest predation in hedgerows and scrub forest in a human-dominated landscape of central Mexico. Acta Oecologica 31, 158–167.

Table 1. Results of the Generalized Linear Mixed Model (binomial distribution with logit link function) analysing the effects on nest predation of distance to edge of tree plantations and height above ground of individual artificial nests located in 30 tree plantations on former cropland (random factor). p: statistical significance was estimated using a robust approach with quasi-ML standard errors. Significant predictor variables at p<0.05 are emboldened. Beta (β): standardized partial regression coefficients; se: standard error of beta.

	df	β	se	p
Distance to plantation edge (m)	1	*-0.207*	0.305	*0.496*
Height above ground (m)	1	*-0.257*	0.223	*0.250*
Plantation (random factor)	**29**			***<<0.001***

Table 2. Results of the Generalized Linear Model (Poisson distribution with log-link function) analyzing the effects of six predictor variables on the number of predated artificial nests at 30 tree plantations on former cropland. The total number of artificial nests placed at each plantation was used as an offset of the model. p: statistical significance was estimated using a robust approach with quasi-ML standard errors. Significant predictor variables at $p<0.05$ are emboldened. Beta (β): standardized partial regression coefficients; se: standard error of beta.

	df	β	se	p
Area (ha; log-transformed)	**1**	***0.127***	**0.055**	***0.021***
Edge/area ratio	**1**	***-0.153***	**0.061**	***0.012***
Magpie abundance (no. individuals)	**1**	***0.169***	**0.061**	***0.006***
PC1 Vegetation structure	**1**	***-0.085***	**0.043**	***0.050***
PC2 Vegetation structure	1	*0.075*	0.041	*0.066*
PC1 Land use types	**1**	***-0.165***	**0.063**	***0.009***

Table 3. Results of the Generalized Linear Mixed Model (binomial distribution with logit link function) showing the effects of predictor variables on predation of artificial nests on open farmland adjacent to 33 tree plantations (random factor). p: statistical significance was estimated using a robust approach with quasi-ML standard errors. Significant predictor variables at $p<0.05$ are bolded. Beta (β): standardized partial regression coefficients for continuous predictors and for the dummy variables built with the levels of the factor habitat types.

	df	β	se	p
Distance to edge (m)	**1**	**-1.122**	**0.363**	**0.002**
Habitat types where nests were placed:	**5**			**0.003**
Olive groves		0.632	0.701	0.368
Vineyard		**8.440**	**0.433**	**<0.001**
Abandoned cropland and pastures		0.656	0.948	0.489
Semi-natural woody vegetation		-0.575	0.650	0.376
Dry herbaceous cropland		0.988	0.800	0.217
Plantation (random factor)	**32**			<<**0.001**

Table 4. Results of the Generalized Linear Model (Poisson distribution with log-link function) analysing the effects of predictor variables on the number of predated artificial nests on open farmland adjacent to 33 tree plantations. The total number of artificial nests in each open farmland habitat was used as an offset of the model. p: statistical significance was estimated using a robust approach with quasi-ML standard errors. Significant predictor variables at $p<0.05$ are bolded. Beta (β): standardized partial regression coefficients; se: standard error of beta.

	df	β	se	p
Plantation area (ha; log-transformed)	**1**	**0.052**	**0.024**	**0.030**
Average tree height in plantations (m)	1	-0.017	0.036	0.641
Magpie abundance (no. individuals)	1	0.016	0.031	0.602
PC1 Land use types	1	-0.061	0.039	0.116

Table 5. Review of nest predation rates at (a) tree plantations, (b) forest fragments and (c) open habitat adjacent to tree plantations or forest fragments. The mean and range of predation rates and the mean ± sd of exposure days for the three habitat types (i.e. a, b and c) have been calculated by the authors of this study on the basis of the referred studies.

Habitat type	Landscape context	Mean (range) predation rate (%)	Mean no. exposed days (±sd)	References
a) Tree plantations		59.5 (23.0-94.0)	12±2	
Conifer plantations	Sub-boreal forest	83.7 (64.7-94)	10	Pedersen et al. 2009
		23.0	14	Vander Haegen & DeGraaf 1996
		41.2 (36.7-45.8)	13	Carignan & Villard 2002
b) Forest fragments		66.4 (38.9-88.0)	11±4	
Forest fragments	Boreal agricultural	88.0	7	Andren 1992
Oak forest fragments	Mediterranean agricultural	87.5	8	Santos & Tellería 1992
		38.9	14	Castilla et al. 2007
Fagus forest fragments	Eurosiberian agricultural	41.7	14	Ludwig et al. 2012
Cloud forest fragments	Andean agricultural	48.9	15	Arango-Vélez & Kattan 1997
Rainforest fragments	Tropical pastures	71.9	9	Estrada et al. 2002
c) Open habitat adjacent to tree plantation or forest fragments		60.0 (13.7-100)	10±3	
Forest fragments	Boreal agricultural	99.0	7	Andren 1992
		41.0	14	Vander Haegen & DeGraaf 1996
Tree plantations and Oak forest fragments	Mediterranean agricultural	49.0	15	Reino et al. 2010

		13.7	14	Castilla et al. 2007
		50.0	8	Santos & Tellería 1992
Fallow	Template forest	86.5	14	Conner & Perkins 2003
Rainforest fragments	Pastures with tropical rainforest remnant	79.0	9	Estrada et al. 2002
Clearing	Turkey Oak forest	24.0	7	Purger et al. 2004

Figure 1. Percentage of predated artificial nests by the first counting and by the second counting and of non predated nests at tree plantations (Experiment 1) and on adjacent open farmland habitat (Experiment 2).

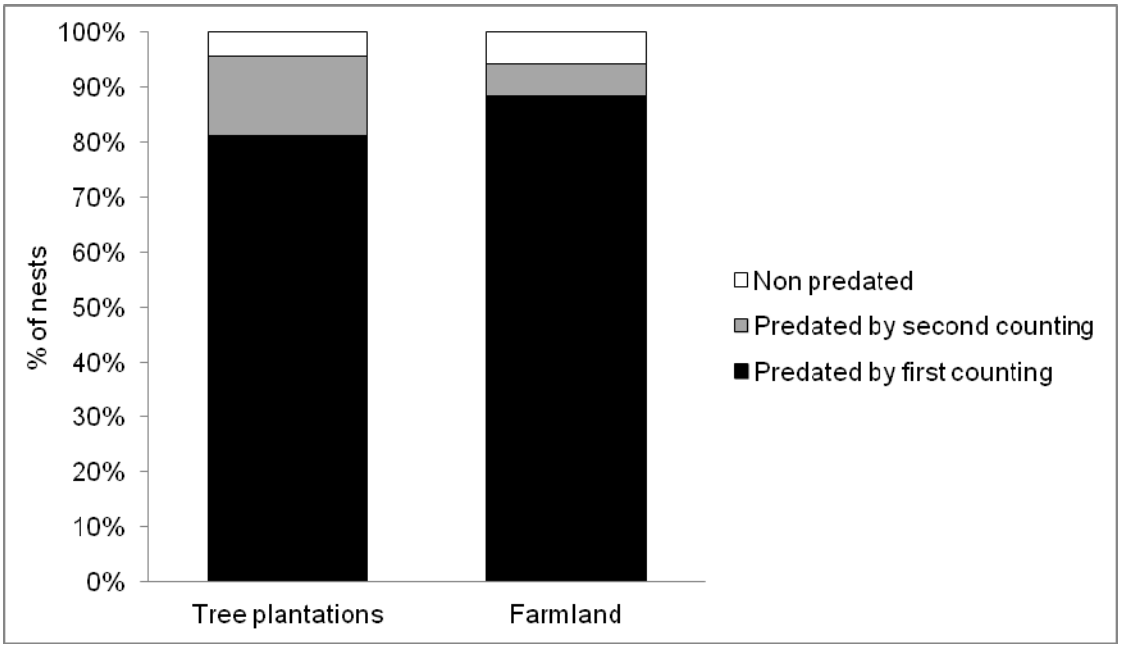

Figure 2. Partial residual plot of the influence of magpie abundance on predation intensity of nests at 30 tree plantations on former cropland. The residual plot shows the relationship with magpie abundance given that the other independent variables are also in the model, therefore partialling out their effects (see **Table 2** for more details).